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# The influence of sulfur dioxide and water on the performance of a marine SCR catalyst

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#### ABSTRACT

This study investigates how sulfur affects the  $NO_x$  reduction activity over a commercial vanadium based urea-SCR catalyst for marine applications, especially at low temperatures, and in combination with  $H_2O$ . The addition of  $SO_2$  in the absence of  $H_2O$  promotes the  $NO_x$  reduction at  $350\,^{\circ}C$ , while the addition of  $H_2O$ , in the absence of  $SO_2$ , gives rise to a decrease in the  $NO_x$  reduction and also an inhibition of the  $N_2O$  formation. The same trends are observed at transient temperatures, but no promotional effect by  $SO_2$  is seen at temperatures below  $230\,^{\circ}C$ . Further, long term effects of  $SO_2$  and  $H_2O$  were investigated and the  $NO_x$  reduction remains stable, also after long term exposure of  $SO_2$ . The ammonia desorption is investigated using temperature programmed desorption (TPD) experiments, both in the presence and in the absence of  $SO_2$ . In general in the presence of both  $H_2O$  and  $SO_2$  the catalyst does not show any sign of deactivation at temperatures above  $300\,^{\circ}C$  and fairly low space velocities (below  $12,200\,h^{-1}$ ). However, at lower temperatures ( $250\,^{\circ}C$ ) and/or higher space velocities the catalytic performance for  $NO_x$  reduction decreases with time.

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#### 1. Introduction

The growing awareness of environmental pollution, such as nitrogen oxides  $(NO_x)$ , causing eutrophication and acidification at land and sea [1], has contributed to the development of more stringent international  $NO_x$  legislations, of which IMO's (International Maritime Organization) Revised MARPOL (Marine Pollution) Annex VI [2] will significantly influence the shipping community. The most stringent  $NO_x$  legislation, Tier III, which will enter into force the 1st of January 2016, has given rise to an increased demand on  $NO_x$  abatement technology for marine applications.

The upcoming Tier III will likely lead to the use of several  $NO_X$  abatement technologies, of which some are not able to reach the Tier III level by themselves. Several different technologies are available on the market [3], such as basic and advanced internal engine modifications (IEM), exhaust gas recirculation (EGR), direct water injection (DWI) and humid air motor (HAM). However, the most effective  $NO_X$  abatement technology on the market is selective catalytic reduction (SCR) using a base-metal catalysts and urea as reducing agent. Already in 1957 it was discovered that ammonia could be used to catalytically remove  $NO_X$  from lean exhaust

gases [4]. The technique was first used to clean exhaust gases from stationary sources such as power plants and chemical plants. During the last 15-20 years ammonia-SCR has been extensively investigated for reduction of  $NO_x$  from heavy-duty vehicles [4–7]. Depending on application there are different possibilities for the ammonia source, e.g. liquefied ammonia, ammonium carbamate, urea dissolved in water and solid urea [7]. In Europe, an infrastructure for the reducing agent (for heavy-duty vehicles) is developed and then the choice is urea dissolved in water [7]. Many catalyst formulations have been investigated for ammonia-SCR and base metal oxides are among the most common commercial ones. The predominant component in such catalysts is vanadium oxide supported on titanium oxide [7], which is the choice for marine applications. Other materials like zeolites have been investigated in this context and for instance Cu-ZSM-5 is found to be active in a broader temperature range compared to vanadium-based catalysts, however, the long-term stability and sulfur resistance may be a problem [7].

The main challenges for SCR in marine applications lies in deactivation and low-temperature activation. The latter is also a problem for cars, trucks and non-road applications and may imply that the catalyst remain inactive during e.g. start-up and maneuvering. The deactivation problem, although SCR catalysts are relative resistant, can be particularly problematic for marine applications where the fuels are residual products that are not well specified, and the urea solution may contain impurities.

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**Table 1**Dimensions of sample A and sample B.

Sample	Cross section area [mm <sup>2</sup> ]	Length [mm]	cpsi	Number of cells
Α	330.5	59.5	25	12
В	2601	34.5	25	100

Further, the low-temperature activity may be further hampered by the presence of impurities (sulfur, salts etc.) in the exhaust. According to Kuiken [8], a typical exhaust gas composition (in vol%) for a two-stroke crosshead engine is 13% O<sub>2</sub>, 75.8% N<sub>2</sub>, 5.2% CO<sub>2</sub>, 5.35% H<sub>2</sub>O, 1500 ppm NO<sub>x</sub>, 600 ppm SO<sub>x</sub>, 60 ppm CO, 180 ppm HC and in addition 120 mg/Nm³ of particulate matter. However, the SO<sub>2</sub> concentration varies with the sulfur content in the fuel used which is in the range 0.1–3 wt.% S, giving rise to an approximate range of 20–650 ppm SO<sub>2</sub> in the exhaust. Furthermore, SCR for marine applications are usually not applied in combination with PM-filters and/or pre-oxidation catalyst. However, sometimes an oxidation catalyst is placed downstream of the SCR in order to reduce the NH<sub>3</sub>-slip and also the CO emissions.

The major part of the scientific research on urea/ammonia-SCR has been devoted to stationary sources and vehicles. The scientific literature on marine applications is sparse, and even though the technique for land-based applications is similar to the marine applications, the boundary conditions are widely differing. However, the impact of sulfur in the exhaust gas on the catalytic activity has been studied for vanadium-based catalysts for stationary sources; see e.g. [9–13], which is also relevant for marine applications. It was, for instance, found that the NO<sub>x</sub> reduction is unaffected and high when either water or SO<sub>2</sub> is present in the feed over a pre-sulfated vanadium-based monolith [10,11]. In the presence of both sulfur dioxide and water, on the other hand, the activity for the NH3-SCR reaction was found to decrease [9,14]. In urea-SCR applications, formation of solid material that can clog pores of the catalyst is a potential problem. For instance, the formation of ammonium sulfate salts has been studied by Huang et al. [12] over vanadiumbased catalysts and it was found that these compounds are formed in the presence of both SO<sub>2</sub> and water at low temperatures (around 250 °C). Furthermore, a few studies have been performed for other poisonous compounds than sulfur. Nicosia et al. [14] studied the effect of urea impurities, like potassium, calcium and phosphate, on the NH<sub>3</sub>-SCR reaction and the deactivation caused by potassium chloride was studied by Zhang et al. [15].

The aim of this study is to investigate how sulfur affects the  $NO_{x}$  reduction activity over a commercial urea-SCR catalyst for marine applications, especially at low temperatures, and in combination with  $H_{2}O$ .

#### 2. Experimental methods

#### 2.1. Catalyst samples

The SCR catalyst samples in the present study were cut from a commercial titanium oxide ( $TiO_2$ )-supported vanadium oxide/tungsten oxide ( $V_2O_5/WO_3$ )-based, washcoated urea-SCR

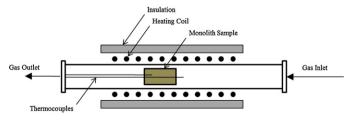


Fig. 1. Laboratory-scale flow reactor system.

 Table 2

 (a-n) Detailed description of the presented experiments

Type of experiment	Sample	SV (h <sup>-1</sup> )	Temperature (°C)	SO <sub>2</sub> (ppm)	H <sub>2</sub> O(%)
Steady-state activity Steady-state activity (a)	A (pre-oxidation for 20 min) 6100, 12,200, 18,300	6100, 12,200, 18,300	250, 300, 350	0	0
Steady-state activity (b)	В	5700	250, 300, 350	0	0
Steady-state activity including $SO_2$ and $H_2O$					
Cyclic variation of SO <sub>2</sub> (c)	A, B	6100, 5700 (respectively)	350	$5 \times (1 \text{ h 0; 0.5 h 100})$	0
Transient ${ m SO}_2\left({ m d} ight)$	A	12,200	350	0.5 h each: 0; 100; 0; 200;0; 1000; 0; 1100	0
Influence of ${\rm CO_2}$ and ${\rm H_2O}\left(e\right)$	А	12,200	350	0	0.5 each 0; 2; 4; 6
Catalytic activity with transient temperature					(0% CO2)
Transient temperature with SO <sub>2</sub> (f)	A	12,200	$400 \rightarrow 100$ and $100 \rightarrow 400$	0; 100; 500; 1000	0
			(10°C/min)		
Transient temperature with $H_2O(g)$	A	12,200	$400 \rightarrow 100$ and $100 \rightarrow 400$	0	4
			(10°C/min)		
Catalytic recovery after SO <sub>2</sub> exposure and long term effects of SO <sub>2</sub> and H <sub>2</sub> O	and H <sub>2</sub> O				
Recovering after SO <sub>2</sub> exposure (h)	A	12,200	350	1 h 500; 6 h 0	0
Long term effect of SO <sub>2</sub> (i)	∢	6100	350	15 h 500	0
Long term effect of $H_2O(j)$	∢	6100	350	0	15 h 4
Catalytic activity with transient temperature and space velocity and simultaneous introduction of SO <sub>2</sub> and H <sub>2</sub> O	d simultaneous introduction of S	30 <sub>2</sub> and H <sub>2</sub> 0			
Long term effect of $SO_2$ and $H_2O$ ; 3 different temperatures (k)	A (pre-oxidation at 350 °C) 6100	6100	250, 300, 350 (5 h each temp)	200	4
Long term effect of $SO_2$ and $H_2O$ ; 2 different SV (1)	A (pre-oxidation at 350 °C)	12,200, 18,300 (5 h each SV) 300	300	200	4
Catalytic activity with transient temperature and SO <sub>2</sub> content (m) A (pre-oxidation at 350 °C)	) A (pre-oxidation at 350 °C)	6100	250, 300, 350 (2 h each temp)	250; 750	4
Temperature programmed desorption (TPD)					
TPD with NH <sub>3</sub> and SO <sub>2</sub> (n)	Α	12,200	44 → 550 °C, (10 °C/min)	0; 500	0
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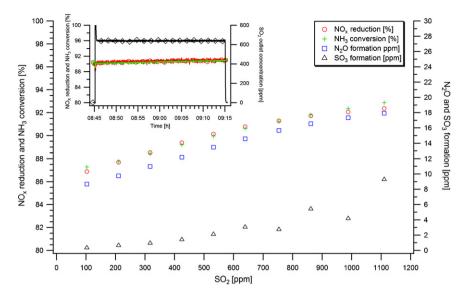


Fig. 2. Influence of the  $SO_2$  concentration on the  $NO_x$  reduction performance over sample A. Feed gas composition: 500 ppm NO, 500 ppm NH<sub>3</sub>, 8%  $O_2$  and 100–1100 ppm  $SO_2$  (350 °C and 12,200 h<sup>-1</sup>).

catalyst for marine applications provided by D.E.C. Marine AB. In order to enable experiments over a range of space velocities (SV) the catalyst was machined into two samples of different sizes (Table 1) (samples A and B).

#### 2.2. Flow reactor experiments

The flow reactor consisted of a horizontal quartz tube heated by a heating coil (Fig. 1), with an outer diameter corresponding to either 25 or 78 mm, for samples A and B, respectively. The inlet and sample temperatures were measured by thermocouples (type K), approximately 15 mm before the sample and inside the monolith sample (just before the rear end), respectively. All stated temperatures in this study are inlet values. The inlet gas composition was controlled by mass flow controllers (Bronkhorst High-Tech) and the outlet gas composition was analyzed by an MKS MultiGas 2030 HS FTIR instrument. For the experiments with both water and  $SO_2$  in the feed gas, a Horiba PG-250 instrument was used to analyze the outlet gas composition. The PG-250 uses non-dispersive IR detection for  $SO_2$ , and  $SO_2$ ; chemiluminescence (cross-flow modulation) for  $SO_2$ , and a galvanic cell for  $SO_2$  measurements.

Prior to all flow reactor experiments the samples were preoxidized in 8% O<sub>2</sub> (Ar bal.) at  $550\,^{\circ}$ C (or  $350\,^{\circ}$ C when the Horiba instrument was used) for 1 h. The inlet gas composition consisted of  $500\,\text{ppm}\,\text{NO}$ ,  $500\,\text{ppm}\,\text{NH}_3$ ,  $8\%\,\text{O}_2$ , with varying concentrations of  $SO_2$  (0–1000 ppm) and  $H_2O$  (0–6%), balanced with Ar. The temperature and SV were also varied. The different experiments are listed in Table 2. Further, temperature programmed desorption (TPD) experiments were performed over sample A in order to evaluate the ammonia desorption in the absence and presence of  $SO_2$ . After pretreatment, NH<sub>3</sub> (500 ppm) was adsorbed (SV 12,200 h<sup>-1</sup> and 44 °C) for 1.5 h. The sample was then purged with Ar for 1.25 h to remove the physically adsorbed NH<sub>3</sub>. A temperature ramp (44  $\rightarrow$  550 °C, 10 °C/min) was then performed in either Ar or in  $SO_2$  (500 ppm) and Ar.

#### 3. Results

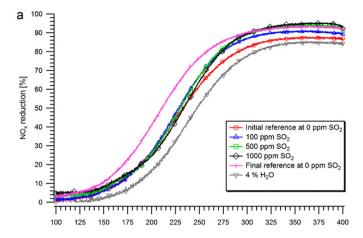
3.1. Steady-state activity for ammonia-SCR with/without  $SO_2$ ,  $H_2O$  and  $CO_2$ 

Steady state measurements (Table 2a and b) were conducted for samples A and B in order to investigate the influence of the space velocity and temperature on the catalytic activity for lean  $NO_x$  reduction with ammonia. The results are summarized in Table 3. It can be seen that the activity decrease with lower temperature and higher space velocity, and in addition it can be noted that sample A (SV 6100 h<sup>-1</sup>) and B (SV 5700 h<sup>-1</sup>) show very similar  $NO_x$  reduction performance. The slight increase over sample B may be attributed to the somewhat lower space velocity.

In order to investigate the potential effect of  $SO_2$  on the catalytic activity, steady state measurements were performed for samples A and B with cyclic variation of  $SO_2$  (Table 2c). The results show stable

Steady state  $NO_x$  reduction over sample A and B. Feed gas composition: 500 ppm NO, 500 ppm  $NH_3$  and  $8\% O_2$ .

Sample	SV [h <sup>-1</sup> ]	Reaction temperature [°C]	NO <sub>x</sub> reduction [%]	NH <sub>3</sub> conversion [%]	N <sub>2</sub> O formation [ppm]
Sample A	6100	350	97.1	98.3	5.5
-	6100	300	97.3	96.8	3.4
	6100	250	93.4	91.3	3.3
Sample A	12,200	350	88.6	90.3	5.9
-	12,200	300	87.0	86.9	3.3
	12,200	250	76.3	75.8	3.1
Sample A	18,300	350	79.2	81.5	6.4
•	18,300	300	76.3	77.0	3.3
	18,300	250	63.5	64.5	2.9
Sample B	5700	350	98.7	95.2	6.1
•	5700	300	98.3	94.1	3.3
	5700	250	93.8	89.2	2.6



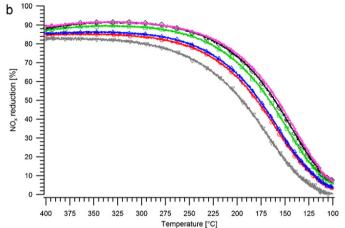


Fig. 3. Catalytic activity for lean  $NO_x$  reduction as a function of temperature (a) 100–400 °C and (b) 400–100 °C).

 $NO_x$  reduction activity in the range of 98–99%, comparable to the catalytic performance without  $SO_2$ .

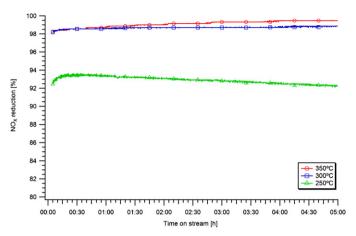
Furthermore, the catalytic activity for  $NO_x$  reduction was studied for increasing  $SO_2$  concentrations (100–1100 ppm) over sample A (Table 2d). The results are shown in Fig. 2, where it can be seen that the SCR activity increases from about 87% to 92% as the  $SO_2$  concentration increases from 100 ppm to 1100 ppm. Also the  $SO_3$  concentration is shown and can be seen to be about 1% of the inlet  $SO_2$  concentration.

The effects of  $H_2O$  and  $CO_2$  on the catalytic activity were also investigated for sample A (Table 2e). The results obtained with  $H_2O$  added to the feed gas are summarized in Table 4. Considering the influence of  $H_2O$  and  $CO_2$  on the  $NO_X$  reduction performance it can be noted that the effect of  $H_2O$  on the catalytic activity is two folded; both a decrease of the  $NO_X$  reduction rate and an inhibition of the  $N_2O$  formation. Further, the effect of  $H_2O$  seems to be fairly independent of the  $H_2O$  concentration.

Finally, the introduction of CO<sub>2</sub> was investigated and did not give rise to any significant changes in the catalytic activity.

## 3.2. Influence of transient temperature on the catalytic activity for $NO_x$ reduction

The influence on the lean  $NO_x$  reduction activity upon transient temperature changes with  $SO_2$  or  $H_2O$  present in the gas feed (Table 2f and g) are presented in Fig. 3. The experiment started with the cooling process, ramping from  $400\,^{\circ}\text{C}$  to  $100\,^{\circ}\text{C}$ , followed by the opposite heating process. This was done for several sulfur concentrations (0, 100, 500, 1000 ppm), starting and finishing with 0 ppm



**Fig. 4.** NO<sub>X</sub> reduction at 350 ( $\bigcirc$ ), 300 ( $\square$ ) and 250 °C ( $\triangle$ ). Feed gas composition: 500 ppm NO, 500 ppm NH<sub>3</sub>, 8% O<sub>2</sub>, 500 ppm SO<sub>2</sub> and 4% H<sub>2</sub>O (SV 6100 h<sup>-1</sup>).

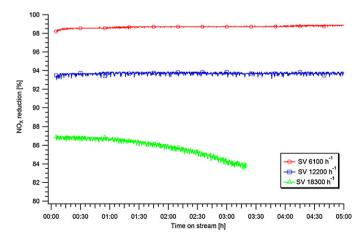
sulfur. The sample was then conditioned (oxidized) in 8% O<sub>2</sub> (Ar bal.) at  $550\,^{\circ}$ C for 1 h, to ensure removal of sulfur compounds from the sample, followed by the equivalent cooling/heating process in the presence of 4% H<sub>2</sub>O. It can be seen that during the cooling process the catalytic activity increases with increasing concentration of SO<sub>2</sub>. During heating, on the other hand, the activity increase, with increasing SO<sub>2</sub> concentration, starts at around  $240\,^{\circ}$ C and becomes higher above  $275\,^{\circ}$ C. In the presence of H<sub>2</sub>O the catalytic activity shows a constant decrease in the whole temperature range.

## 3.3. Long term effects of $SO_2$ and $H_2O$ on the catalytic $NO_X$ reduction

The prolonged effect of  $SO_2$  exposure was investigated over sample A (Table 2h). During an hour, in the presence of  $SO_2$ , the  $NO_x$  reduction increased slightly (from 85% to 88%). This increase in activity was then sustained during the following 6 h in the absence of  $SO_2$ 

When continuously adding  $SO_2$  or  $H_2O$ , separately, to the gas feed (Table 2i and j) over sample A, the lean  $NO_x$  reduction performance is stable, corresponding to 99 and 97% for  $SO_2$  and  $H_2O$ , respectively.

The long term influence on the catalytic activity for  $NO_x$  reduction was also investigated with  $SO_2$  and  $H_2O$  added to the gas feed simultaneously (Table 2k-m). The influences of temperature, space velocity and  $SO_2$  are presented in Figs. 4–6, respectively. It was found that the catalytic activity remains high (98%) and



**Fig. 5.** NO<sub>x</sub> reduction for 6100 ( $\bigcirc$ ), 12200 ( $\square$ ) and 18,300 h<sup>-1</sup> ( $\triangle$ ). Feed gas composition: 500 ppm NO, 500 ppm NH<sub>3</sub>, 8% O<sub>2</sub>, 500 ppm SO<sub>2</sub> and 4% H<sub>2</sub>O (300  $^{\circ}$ C).

**Table 4**Influence of H<sub>2</sub>O and CO<sub>2</sub> on the NO<sub>3</sub> reduction performance. Feed gas composition: 500 ppm NO, 500 ppm NH<sub>3</sub>, 8% O<sub>2</sub>, 0–6% H<sub>2</sub>O (350 °C and 12,200 h<sup>-1</sup>).

Step [0.5 h/step]	H <sub>2</sub> O [%]	NO <sub>x</sub> reduction [%]	NH <sub>3</sub> conversion [%]	N <sub>2</sub> O formation [ppm]
Step 1	0	86.0	87.5	8.5
Step 2	2	82.7	89.4	1.1
Step 3	0	86.7	88.4	8.9
Step 4	4	82.4	84.1	1.1
Step 5	0	87.1	88.2	9.1
Step 6	6	82.2	82.9	1.2
Step 7	0	87.4	88.3	9.2

stable at temperatures above 300 °C, while at 250 °C showing an instant decrease in catalytic activity (93%) followed by a continuous decrease (Fig. 4). Further, changing the space velocity results in a direct effect on the catalytic activity (Fig. 5). The two lower space velocities (6100 and 12,200 h $^{-1}$ ) allow stable activity, while the higher space velocity (18,300 h $^{-1}$ ) gives rise to a continuous decrease in catalytic activity. The final experiment, investigating the effect of sulfur concentrations at different temperatures, shows high and stable catalytic activities at 350 °C and 300 °C, while a significant drop in the catalytic activity is observed at 250 °C for both 250 ppm and 750 ppm SO<sub>2</sub> (Fig. 6). At 250 °C and 750 ppm SO<sub>2</sub> a continuous decrease in activity is observed.

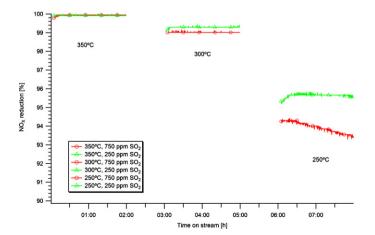
#### 3.4. Temperature programmed desorption

In order to evaluate the influence of  $SO_2$  on the ammonia desorption, temperature programmed desorption (TPD) experiments were conducted (Table 2n). The results are illustrated in Fig. 7.

The TPD curves show a maximum in  $NH_3$  desorption around  $150\,^{\circ}\text{C}$  and a shoulder above  $225\,^{\circ}\text{C}$  in both presence and absence of  $SO_2$ . This indicates two types of adsorption sites for  $NH_3$  on the sample surface, one weaker and one stronger interaction with  $NH_3$ . At  $150\,^{\circ}\text{C}$  the  $NH_3$  desorption is higher without  $SO_2$  in the feed, while the desorption peak around  $275\,^{\circ}\text{C}$  is slightly higher in the presence of  $SO_2$ . In total the amount of adsorbed ammonia is slightly higher in the absence of  $SO_2$  (about 3.5%).

#### 4. Discussion

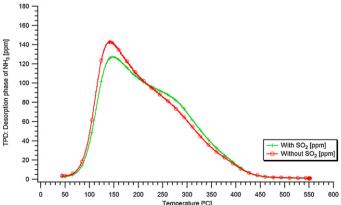
Considering the results for sample A it is evident that the  $NO_x$  reduction rate is dependent both on the space velocity and on the reaction temperature (Table 3). This is illustrated by comparing the  $NO_x$  reduction at SV 18,300 h<sup>-1</sup>/300 °C and SV 12,200 h<sup>-1</sup>/250 °C; of which both are in the range of 76%. Further, the steady state



**Fig. 6.** NO $_x$  reduction at 350, 300 and 250  $^{\circ}$ C for 250 ( $\triangle$ ) and 750 ppm SO $_2$  ( $\bigcirc$ ). Feed gas composition: 500 ppm NO, 500 ppm NH $_3$ , 8% O $_2$ , 250 and 750 ppm SO $_2$  and 4% H $_2$ O (SV 6100 h $^{-1}$ ).

 $NO_x$  reduction for sample B is in the range of 94–99% with a rather weak dependence on reaction temperature, which may be ascribed to the low space velocity (5700 h<sup>-1</sup>) (Table 3). This is similar to the  $NO_x$  reduction activity for the sample A measurements at SV 6100 h<sup>-1</sup> (93–97%). A slight increase in  $NO_x$  reduction is noted for sample B in comparison to sample A (at SV = 6100 h<sup>-1</sup>), which may be explained by the lower space velocity (5700 h<sup>-1</sup>) for sample B. Hence, the comparable results for samples A and B at low space velocity, renders the use of a small sample (sample A) possible, which widens the reaction conditions that can be used. However, sample A only contains 12 cells, compared to 100 cells for sample B, implying that even a small variation in, *e.g.* the wash coat composition may influence the result for sample A to a larger extent than for sample B.

The catalytic activity for  $NO_x$  reduction with increasing  $SO_2$ concentrations (100-1100 ppm) over sample A is increasing with increasing concentration of SO2 in the reaction gas mixture together with increased N2O formation (Fig. 2). These results are supported by earlier studies regarding the influence of SO<sub>2</sub> on lean NO<sub>x</sub> reduction by ammonia over a V<sub>2</sub>O<sub>5</sub>/AC (activated carbon) catalyst at low temperatures (180-250 °C) [16-18]. According to Zhu et al. [17] it is suggested that the SO<sub>2</sub> is adsorbed and oxidized to SO<sub>3</sub> on the vanadium surface and then migrates to the carbon surface where it converts to sulfate species through reaction with H<sub>2</sub>O. The formation of SO<sub>3</sub> is confirmed by the data in Fig. 2 and also suggested by Orsenigo et al. [19] as a first step to formation of sulfate species. The formed sulfate species act as new acid sites, which improves the NH<sub>3</sub> adsorption and thereby enhances the SCR activity. During the SCR reaction, in the presence of SO<sub>2</sub> at a temperature at around 180-250 °C, the formed sulfate species stay on the catalyst surface, while the ammonium ions react with NO continuously to avoid the formation and deposition of excess ammonium sulfate salts on the catalyst surface resulting in the catalyst being promoted but not poisoned by SO<sub>2</sub>. Further, Zhu et al. [18] found that the promoting effect of SO<sub>2</sub> on a V<sub>2</sub>O<sub>5</sub>/AC (Sn-doped) is attributed to the



**Fig. 7.** Influence of  $SO_2$  on the desorption of  $NH_3$ , using TPD. Feed gas composition: 500 ppm  $NH_3$  during adsorption, desorption during temperature ramp with/without 500 ppm  $SO_2$  (12,200  $h^{-1}$ ).

formation of sulfate species on the catalyst surface, which provides new Brønsted acid sites that increases the ammonium adsorption and the catalytic activity. However, the effect of the formed sulfate species is suggested to be two folded: (1) it provides acid sites for ammonium adsorption and thus enhancing the catalytic activity, (2) the sulfate species react with ammonia and transform into ammonium sulfate salts, such as NH<sub>4</sub>HSO<sub>4</sub> and (NH<sub>4</sub>)<sub>2</sub>S<sub>2</sub>O<sub>7</sub>, which may plug the catalyst pores resulting in catalytic deactivation. It was also found by Huang et al. [16] that SO<sub>2</sub> promotes the SCR activity, over a V<sub>2</sub>O<sub>5</sub>/AC catalyst, in the absence of H<sub>2</sub>O, while in the presence of H<sub>2</sub>O it speeds the deactivation. The two folded effect of  $SO_2$  is attributed to the  $SO_4{}^{2-}$  formed on the catalyst surface which stays as ammonium sulfate salts on the surface. In the absence of H<sub>2</sub>O a small amount of the salts are deposited on the catalyst surface resulting in a promoted SCR activity, while in the presence of H<sub>2</sub>O the deposition rate of ammonium sulfate salts becomes much higher resulting in pore blocking and related catalytic deactivation.

The SCR reaction mechanism over a V<sub>2</sub>O5/TiO<sub>2</sub> catalyst typically involves two types of sites, where NH3 adsorbs on a Brønsted acid site (V-OH), which in combination with a shift in oxidation state of a V=O site catalyze the reduction of gas-phase NO [20]. Guo et al. [21] studied the effect of sulfate species on a V<sub>2</sub>O<sub>5</sub>/TiO<sub>2</sub> catalyst and proposed that  $SO_2$  is oxidized to  $SO_3$  over  $V_2O_x$  sites, which subsequently forms sulfate species on TiO<sub>2</sub>. They state that vanadium and hydroxyl ions are highly mobile under high-temperature reaction conditions (350-400 °C) in the presence of water, which leads to migration of SO<sub>3</sub> from a vanadium site to a neighboring Ti-OH site forming a S-OH group. This S-OH group can also act as a Brønsted acid site, which increases the ammonia adsorption capacity at these conditions. Further, Kijlstra et al. [22] investigated the promotion and deactivation of a V<sub>2</sub>O<sub>5</sub>/TiO<sub>2</sub> SCR catalyst by SO<sub>2</sub> at low temperature, below 200 °C, and it was found that at lower V<sub>2</sub>O<sub>5</sub> loadings, the surface of the catalysts can be sulfated by SO<sub>2</sub>, which gives rise to enhanced Brønsted acidity and to a higher SCR activity. On the other hand, at higher V<sub>2</sub>O<sub>5</sub> loadings (6 wt.%) the SO<sub>2</sub> oxidation increases, followed by increased ammonium sulfate (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub>, formation and deposition, which overrules the promoting effect and leads to deactivation of the catalyst.

The observed promoting effect by  $SO_2$  is thus suggested to be attributed to the formation of new Brønsted acid sites (S-OH) that enhance the  $NH_3$  adsorption and thereby promote the  $NH_3$  reaction with NO.

The influence of  $H_2O$  has been observed earlier [16,23] over  $V_2O_5/AC$  catalysts and the results indicate an inhibiting effect on the NO reduction, which is suggested to be caused by competitive adsorption of  $H_2O$  and the reactants, such as  $NH_3$  and/or NO on vanadia sites. It is also suggested that the effect of  $H_2O$  is independent of the vanadia loading. The influence of  $H_2O$  has also been investigated on other types of catalysts targeting selective catalytic reduction of NO, such as modified activated carbon fibres [24] and  $MNO_x/Al_2O_3$  [25]. It was suggested that  $H_2O$  inhibits NO reduction at low temperatures (below  $250\,^{\circ}C$ ) due to competitive adsorption. However, even in the presence of water an excess of adsorbed ammonia was observed on the catalyst surface [25] and the decreased SCR reaction rate is considered to be caused by decreased adsorption of weakly bound nitric oxide.

Further, the inhibiting effect of  $H_2O$  on the  $N_2O$  formation, supported by, *e.g.* [25,26]. Kijlstra et al. [25], show a reversible and positive effect on the  $N_2$  selectivity for  $Mn/Al_2O_3$  at 325 °C and  $Mn_2O_3$  at 150 °C, where the inhibition of ammonia oxidation and the non-selective catalytic reduction of nitric oxide are suggested to be stronger than the inhibition of the SCR reaction. According to Bagnasco et al. [26], who studied the selective reduction of NO with  $NH_3$  over an iron-vanadyl phosphate catalyst, the addition of 1000 ppm  $H_2O$  (at 200–400 °C) results in a decrease in  $N_2O$  formation, thus enhancing the reaction selectivity. This behavior is

ascribed to an inhibiting effect of water on the  $NH_3$  oxidation. Hence, despite the different types of SCR catalysts studied, these results support the observed inhibition effect by  $H_2O$  on the  $N_2O$  formation reported in this work.

The addition of  $H_2O$ , in absence of  $SO_2$ , at a reaction temperature of  $350\,^{\circ}C$  and SV of  $12000\,h^{-1}$  is observed to have a twofold effect on the SCR activity; inhibition of the  $NO_X$  reduction (from about 87% down to 82%) and inhibition of the  $N_2O$  formation (from about 9 ppm down to 1 ppm). The effect of  $H_2O$  seems to be fairly independent of the  $H_2O$  concentration (2–6%). The decreased  $NO_X$  reduction by  $H_2O$  is suggested to be attributed to competitive adsorption of  $H_2O$  on vanadia sites and the reactants such as  $NH_3$  and/or NO, while the inhibition of the  $N_2O$  formation may be attributed to reduced activity for ammonia oxidation.

The prolonged effect of  $SO_2$  exposure was investigated over sample A (Table 2h) and the  $NO_X$  reduction remained stable for 6 h after 1 h of  $SO_2$  exposure (350 °C and SV 6100 h<sup>-1</sup>). This eliminates any exothermic effects, causing the enhancement, that may have occurred due to  $SO_2$  being oxidized to  $SO_3$ . Further, this is consistent with a stable promoting effect of  $SO_2$  and that  $NH_3SO_4$  is unstable at this temperature.

Further, when continuously adding  $H_2O$ , separately, to the gas feed (Table 2j) over sample A, the lean  $NO_x$  reduction performance is stable at 97%. However, the effect of  $H_2O$  seems to be very dependent on the space velocity, e.g. comparing the results in Table 4 (SV  $12,200\,h^{-1}$ ) with a  $NO_x$  reduction of approximately 82%. Thus, in the presence of  $H_2O$  the  $NO_x$  reduction seems even more sensitive to space velocity than in the absence of water. This is in agreement with competitive adsorption of  $H_2O$  and the reactants.

The results presented in Fig. 3 indicate a promoting effect of  $SO_2$  on the  $NO_x$  reduction, higher for increasing  $SO_2$  concentrations (Table 2f). The catalytic activity during heating is observed to increase with increased  $SO_2$  concentration (at temperatures above  $230\,^{\circ}\text{C}$ ), while more pronounced during cooling.

At lower temperatures the catalytic activity is similar for all SO<sub>2</sub> concentrations. During the cooling process there is an obvious promotional effect from SO<sub>2</sub>, in the whole temperature range. This effect seems to be permanent since the higher activity seen in the presence of SO<sub>2</sub> is sustained also after removal of SO<sub>2</sub> from the feed gas. The presence of H<sub>2</sub>O gives reduced activity during both heating and cooling. These observations are consistent with a scenario where the addition of SO<sub>2</sub> to the catalyst has two different effects; the formation of stable Brønsted acid sites that enhance the activity, and the formation of ammonium sulfate that inhibit the SCR activity. These salts seem to decompose at around 230 °C and explain the higher activity with SO<sub>2</sub> at higher temperature in the heating ramps. This decomposition is observed as a shoulder in the TPD curve presented in Fig. 7. Overall the TPD results suggest the formation of ammonium sulfate salts at lower temperatures and start of decomposition at an approximate reaction temperature of 230 °C. In the presence of SO<sub>2</sub> the stored ammonia likely reacts to form ammonium sulfate salts, which needs higher temperature to decompose and desorb than NH<sub>3</sub>. Though, the total desorbed NH<sub>3</sub> is higher in the absence of  $SO_2$ .

During the cooling ramps the ammonium sulfate has not been formed and only the promoting effect of  $SO_2$  is observed. Further, comparing the two experiments without  $SO_2$  (at the end of each ramp series) one can conclude that there is a promoting effect remaining even without  $SO_2$  present, while the ammonium sulfate has decomposed. Finally, the lower  $NO_X$  reduction in the presence of  $H_2O$  is in agreement with competitive adsorption of  $H_2O$  and the reactants.

Considering the influence of temperature on the  $NO_x$  reduction, in the presence of 500 ppm  $SO_2$  and 4%  $H_2O$ , and at low space velocity (6100  $h^{-1}$ ) the results show a slight increase of the catalytic activity over time at 350 °C, constant reduction over time

at 300 °C and a continuous decrease with time at 250 °C (Fig. 4). These results can be compared with the steady state  $NO_x$  reduction without SO<sub>2</sub> and H<sub>2</sub>O, Table 3, where the NO<sub>x</sub> reduction is in the range of 97% at both 300 °C and 350 °C. This implies that the effects of SO<sub>2</sub> and H<sub>2</sub>O are minor at low space velocities at temperatures above 300 °C. One possible explanation to the decrease in activity at 250 °C is the formation and deposition of ammonium sulfate salts, which is consistent with the previous discussion. This has also been reported for V<sub>2</sub>O<sub>5</sub>/AC catalysts exposed to SO<sub>2</sub> and H<sub>2</sub>O [12,16]. According to Huang et al. [12] a higher content of  $H_2O$  (2.5–20%) results in a higher deactivation rate of the catalyst. The addition of H<sub>2</sub>O increases the formation rate of ammonium sulfate salts and decreases the reaction rate between the ammonium sulfate salts and NO, while the SO<sub>2</sub> does not affect the reaction rate. Further, according to Huang et al. [16] a higher V<sub>2</sub>O<sub>5</sub> loading increases the oxidation activity of the catalyst and gives a higher formation rate of the ammonium sulfate salts. It was also found that a higher space velocity results in higher formation rate of ammonium sulfate salts, whereby it is suggested that the catalytic deactivation can be reduced or eliminated by reduced V<sub>2</sub>O<sub>5</sub>-loading and/or reduced space velocity.

Overall, the present results suggest that at a space velocity of  $6100\,h^{-1}$  and temperatures above  $300\,^{\circ}\text{C}$  the  $NO_x$  reduction remain stable, even in the presence of combined  $SO_2$  and  $H_2O$ , and is also comparable to steady state conditions in the absence of both  $SO_2$  and  $H_2O$ . However, at  $250\,^{\circ}\text{C}$  it is suggested that an increased deposition of ammonium sulfate salts gives rise to a decrease of the  $NO_x$  reduction, possibly by pore plunging.

The long term experiments show that there is a continuous deactivation at low temperatures (Fig. 4), high space velocities (Fig. 5) and at high  $SO_2$  concentrations (Fig. 6). This is all consistent with the formation of ammonium sulfates that inhibits the SCR reaction.

#### 5. Conclusions

This paper reports on the effects of  $SO_2$  and  $H_2O$  concentrations, reaction temperature and space velocity on the SCR activity for a commercial vanadia-based washcoated urea-SCR catalyst for marine applications.

In general, addition of  $SO_2$ , in absence of water, promotes the  $NO_{\chi}$  reduction and the  $NH_3$  conversion along with an increased formation of  $N_2O$ . These effects seem to be more pronounced with increasing  $SO_2$  concentration and relatively independent of temperature. The promoting effect on the  $NO_{\chi}$  reduction is sustained also after removing  $SO_2$  from the feed gas. On the contrary, the addition of  $H_2O$ , in absence of  $SO_2$ , results in a decreased  $NO_{\chi}$  reduction and an inhibition of the  $N_2O$  formation. These effects are reversible and fairly independent of both water content and temperature.

In the presence of both  $SO_2$  and water the activity for  $NO_X$  reduction decreases, in particular at high concentrations of  $SO_2$ , low temperatures and high space velocities. This can possibly be explained by the formation of ammonium sulfate salts at low temperatures, which is also consistent with the results from the TPD experiments where, in the presence of  $SO_2$ , a higher desorption of ammonia is noted at the decomposition temperature of ammonium sulfate salts. In conclusion, these problems can thus be avoided by keeping the temperature high enough (i.e. above  $300\,^{\circ}C$ ) and the

space velocity and the  $SO_2$  concentration low enough (i.e. below  $12,200\,h^{-1}$  and  $500\,ppm$ , respectively).

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#### References

- [1] A. Adams, M. Ajero, M. Amann, C. Andersson, A. Barkman, T. Bellander, E. Dame, J. Dawson, L. Emberson, M. Engardt, A. Engleryd, R. Fernandez, P. Grennfelt, H.C. Hansson, K. Hicks, T. Hu, C. Huizenga, M. Iyngararasan, P.E. Karlsson, T. Keating, T. Kjellstrom, J. Klingberg, J. Kuylenstierna, R. Maas, R. Mathur, F. Moldan, J. Moldanová, J. Munthe, J. Pershagen, S. Punte, S. Sethi, S. Shrestha, D. Simpson, D. Winner, Air Pollution & Climate Change, Two Sides of the Same Coin? Swedish Environmental Protection Agency and the Authors, Solna, Sweden, 2009.
- [2] International Maritime Organization, Revised MARPOL Annex VI, Regulations for the Prevention of Air Pollution from Ships and NO<sub>x</sub> Technical Code 2008, 2009 edition, International Maritime Organization, London, United Kingdom, 2009.
- [3] E.D. Jonge, C. Hugi, D. Cooper, Entec UK Limited, Service Contract on Ship Emissions: Assignment, Abatement and Market-based Instruments, Task 2b NO<sub>x</sub> Abatement, Final Report ed., European Commission Directorate General Environment, Nortwich, England, 2005.
- [4] R.M. Heck, R.J. Farrauto, Catalytic Air Pollution Control Commercial Technology, Van Nostrand Reinhold, New York, USA, 1995.
- [5] V.I. Parvulescu, P. Grange, B. Delmon, Catalysis Today 46 (1998) 233.
- [6] M. Koebel, M. Elsener, M. Kleemann, Catalysis Today 59 (2000) 335.
- [7] P.L.T. Gabrielsson, Topics in Catalysis 28 (2004) 177–184.
- [8] K. Kuiken, Diesel Engines Part II for Ship Propulsion and Power Plants, from 0 to 100,000 kW, Target Global Energy Training, Onnen, The Netherlands, 2008.
- [9] H.H. Phil, M.P. Reddy, P.A. Kumar, L.K. Ju, J.S. Hyo, Applied Catalysis B: Environmental 78 (2008) 301–308.
- [10] E. Garcia-Bordeje, J.L. Pinilla, M.J. Lazaro, R. Moliner, Applied Catalysis B: Environmental 66 (2006) 281–287.
- [11] E. Garcia-Bordeje, J.L. Pinilla, M.J. Lazaro, R. Moliner, J.L.G. Fierro, Journal of Catalysis 233 (2005) 166–175.
- [12] Z.G. Huang, Z.P. Zhu, Z.Y. Liu, Q.Y. Liu, Journal of Catalysis 214 (2003) 213–219.
- [13] Z.P. Zhu, H.X. Niu, Z.Y. Liu, S.J. Liu, Journal of Catalysis 195 (2000) 268–278.
- [14] D. Nicosia, M. Elsener, O. Krocher, P. Jansohn, Topics in Catalysis 42–43 (2007) 333–336.
- [15] X.L. Zhang, Z.G. Huang, Z.Y. Liu, Catalysis Communications 9 (2008) 842-846.
- [16] Z.G. Huang, Z.P. Zhu, Z.Y. Liu, Applied Catalysis B: Environmental 39 (2002) 361–368.
- [17] Z.P. Zhu, Z.Y. Liu, H.X. Niu, S.J. Liu, T.D. Hu, T. Liu, Y.N. Xie, Journal of Catalysis 197 (2001) 6–16.
- [18] Z.P. Zhu, Z.Y. Liu, S.J. Liu, H.X. Niu, Applied Catalysis B: Environmental 30 (2001) 267–276.
- [19] C. Orsenigo, L. Lietti, E. Tronconi, P. Forzatti, F. Bregani, Industrial and Engineering Chemistry Research 37 (1998) 2350–2359.
- [20] N.Y. Topsoe, J.A. Dumesic, H. Topsoe, Journal of Catalysis 151 (1995) 241–252.
- [21] X.Y. Guo, C. Bartholomew, W. Hecker, L.L. Baxter, Applied Catalysis B: Environmental 92 (2009) 30–40.
- [22] W.S. Kijlstra, N.J. Komen, A. Andreini, E.K. Poels, A. Bliek, 11th International Congress on Catalysis – 40th Anniversary Pts A and B, vol. 101, 1996, pp. 951–960.
- [23] M.D. Amiridis, I.E. Wachs, G. Deo, J.M. Jehng, D.S. Kim, Journal of Catalysis 161 (1996) 247–253.
- [24] J. Muniz, G. Marban, A.B. Fuertes, Applied Catalysis B: Environmental 27 (2000) 27–36.
- [25] W.S. Kijlstra, J. Daamen, J.M. vandeGraaf, B. vanderLinden, E.K. Poels, A. Bliek, Applied Catalysis B: Environmental 7 (1996) 337–357.
- [26] G. Bagnasco, G. Busca, P. Galli, M.A. Massucci, K. Melanova, P. Patrono, G. Ramis, M. Turco, Applied Catalysis B: Environmental 28 (2000) 135–142.